ANNUAL ENVIRONMENTAL MONITORING REPORT

JANUARY - DECEMBER 1977

HEALTH PHYSICS STAFF

PLANT ENGINEERING DEPARTMENT

STANFORD LINEAR ACCELERATOR CENTER

STANFORD UNIVERSITY

Stanford, California 94305

PREPARED FOR THE

DEPARTMENT OF ENERGY

UNDER CONTRACT NO. EY-76-C-03-5151

May 1978

CONTENTS

INTRODUCTION

The Stanford Linear Accelerator Center (SLAC) is located two miles west of the Stanford campus in San Mateo County, California. Its boundaries include Sand Hill Road on the north, and San Francisquito Creek on the south. The land slopes to the south toward San Francisquito Creek. The total length of the accelerator and experimental areas is approximately 3 miles, and is oriented almost east-west. Figures 1 and 2 locate SLAC with respect to the surrounding vicinity.

SLAC is a large research laboratory devoted to theoretical and experimental research in high energy physics and to the development of new techniques in high energy accelerator particle detectors. The main tool of the laboratory is a 2 mile long linear accelerator. This accelerator produces beams of electrons with energies up to 22 billion electron volts (22 GeV). It can also accelerate positrons, the "antiparticles" of the electrons, up to 15 GeV. The work is carried out under the sponsorship and financial support of the Department of Energy.

Authorization of the project was given by the U.S. Congress in 1961. Construction of the accelerator started in 1962, and was completed in 1966. Research consisting of numerous and varied experiments has been under way since late 1966.

Summary

Environmental monitoring results continue to demonstrate that, except for penetrating radiation, environmental radiological impact due to SLAC operation is not distinguishable from natural environmental sources. During 1977, the maximum neutron dose near the site boundary

was 8.2 mrem. This represents about 8.2% of the annual dose from
natural sources at this elevation, and 1.6% of the technical
standard of 500 mrem per person annually.[1]

There have been no measurable increases in radioactivity in
ground water attributable to SLAC operations. Because of the extended
drought, very few well water samples were collected and analyzed
during 1977. Construction activities have temporarily placed our
sampling stations for the sanitary and storm sewers out of service.
They will be re-established as soon as construction activities permit.
Airborne radioactivity released from SLAC also continues to make only
a negligible entironmental impact, and results in a site boundary
annual dose of less than 0.01 mrem; this represents less than 0.01%
of the annual dose from the natural radiation environment, and about
0.002% of the technical standard.

MONITORING TECHNIQUES AND STANDARDS

Concentration Guides for Liquid Effluent

Because of the nature of the radionuclides produced at SLAC,
the appropriate Concentration Guide (CG) for liquid effluents is
3×10^{-6} μCi/mℓ. This is true because the following isotopes are not
produced at SLAC: ^{90}Sr, ^{125}I, ^{126}I, ^{129}I, ^{131}I, ^{210}Pb, ^{210}Po, ^{211}At,
^{223}Ra, ^{224}Ra, ^{226}Ra, ^{227}Ac, ^{228}Ra, ^{230}Th, ^{231}Pa, ^{232}Th, ^{248}Cm, ^{254}Cf,
^{256}Fm, natural thorium.[1]

Sample Analysis

All water supplies collected are analyzed for gross beta activity and tritium by an independent laboratory. Analysis for alpha-emitting radioelements is not performed because we do not produce nor possess in unsealed form significant quantitites of this type of radioactivity at SLAC.[2] In addition to routine gross beta analysis, SLAC examines aliquots of selected specimens by gamma spectroscopy if necessary.

Sample Preparation

When environmental water samples exhibit large amounts of insoluble solids, the solid fraction and filtrate activities are determined separately.

The total volume of each sample is measured and filtered. The original sample bottle is rinsed with the filtrate, and the activity in each fraction is determined separately.

For the solid fraction, the filter paper and solid are dried at ~100°C and weighed after cooling. A 300 to 500 mg aliquot is counted.

The filtrate and water samples exhibiting low solid content are treated the same. A 200 cc volume is taken from each 500 cc sample and evaporated to a smaller volume in a beaker. It is then transferred to a weighed planchet with dilute acid and water rinses. The sample is evaporated gently, weighed, and counted.

Gross Beta

The prepared samples are counted in a low-background proportional counter ($\leq$ 1.5 cpm). The results are normalized by comparison with the count of a known ^{137}Cs source.

Some samples require a correction for naturally occurring ^{40}K. In that case, the total amount of potassium in the sample is determined by flame photometry. The amount of ^{40}K activity can then be calculated, based on the percent abundance of this naturally occurring radioelement and subtracted from the sample count. This correction is performed on water samples whose gross beta activity is greater than 1×10^{-7} µCi/mℓ.

Tritium

To determine the tritium concentration, a 2-3 mℓ aliquot of the filtered solution from the gross beta sample preparation is accurately transferred to a liquid scintillation sample vial. This sample vial is placed in a liquid scintillation counting system. The limit of sensitivity at the 95% confidence level is $\geq 3 \times 10^{-6}$ µCi/mℓ.

Concentration Guides for Airborne Radioactivity

The Concentration Guides (CGs) for airborne radioactivity appear in Reference 1. They were derived from dose standards which require that no individual in the general population be exposed to greater than 500 mrem in one year.

Airborne radioactivity produced as the result of operations is short-lived; i.e., the half-lives range from 2 minutes to 1.8 hours, and are in gaseous (not particulate) form. These isotopes include the following:

TABLE 1

GASEOUS RADIOACTIVITY RELEASED TO ATMOSPHERE

Isotope	Half-Life	CG μCi/mℓ
^{15}O	2.1 min	5×10^{-8} (a)
^{13}N	9.9 min	5×10^{-8} (a)
^{11}C	20.5 min	5×10^{-8} (a)
^{41}Ar	1.8 hr.	4×10^{-8}

(a) Calculated from Reference 3, assuming total submersion.

Since we do not routinely release airborne radioactivity while

the beam is on, and require a waiting period before turning on the

exhaustors, the only radioisotope released is ^{41}Ar. By far the greater

proportion of exposure an individual may receive, under any circumstances,

from the radioelements listed in Talbe 1 is from whole body immersion.

Thus, for an individual to receive a whole body dose of 500 mrem annually

requires a continuous exposure to a large cloud of ^{41}Ar whose average

concentration equals 4×10^{-8} μCi/mℓ(Ci/m^3) for an entire year.

Analysis Techniques for Airborne Radioactivity

The accelerator and beam switchyard (BSY) areas are vented by

20 fans: the discharge point is just slightly above roof elevation.

The total exhaust rate for the accelerator is 60 m^3/s, and the BSY is

$40 \text{ m}^3/\text{s}$. The accelerator and BSY are not normally vented while the electron beam is on. If personnel entry has to be made during an operating cycle, the area is vented for 10 minutes prior to entry and after the electron beam has been shut off. The release of radioactivity is, therefore, infrequent, and only for brief periods of from 30-60 minutes. The accelerator does not represent a measurable source of gaseous or particulate radioactivity due to low activating potential.

Each BSY ventilation fan is interlocked with a radioactive gas detector comprised of a Geiger-Mueller detector, power supply, rate meter, strip chart recorder, and air pump. The electronics are in continuous operation, and the recorder and air pump are interlocked with the ventilation fan so that they operate only when the machine is being vented.

The gas monitors for the BSY collect particulate samples during venting and have revealed negative results. During this period, no particulate radioactivity above background was detected. This agrees with previous "grab" samples collected in the exhaust stream.

Penetrating Radiation Monitoring Techniques

Seven Peripheral Monitoring Stations (PMS) serve to provide continuously recorded data concerning radiation levels (γ and n) near SLAC boundaries. Their positions are located in Figures 3 and 4.

Radiation information is obtained with a Geiger tube for gamma detection, and a paraffin-moderated BF_3 neutron detector calibrated with a Pu Be neutron source. The resultant sensitivities are such that a γ flux from ^{60}Co exposed to an average value of 1 mR/hr for one hour

will be recorded as ~10^4 counts on the Geiger tube channel, and a neutron flux having an average value for one hour of 1 mrem/hr will be recorded as ~10^5 counts on the BF_3 channel. The hourly printout cycle of the Sodeco register is programmed by two clock motors with cam actuated switches and associated electronic circuitry. This programmer automatically interrupts data acquisition, generates a print command, resets the digits in parallel, and reverts to the normal condition of serial counting of incoming data pulses. Dead time per printout cycle is less than 20 seconds per hour.

In connection with the pulse pair resolution limitation of the Sodeco register mentioned above, an important feature of this system involves the pulser which drives the register. It is of the nonparalyzable type. This means that if the instantaneous counting rate (20 counts/sec) is ever exceeded, the register will merely not count the pulses in excess of its maximum rate. It can count at this maximum rate continuously (72,000 counts/hr).

MONITORING RESULTS

Penetrating Radiation Monitoring

The measured annual dose to the general population from accelerator operations is almost entirely from fast neutrons and is characterized as skyshine from SLAC's research area. The total dose measured at each PMS location appears in Table 2. Estimates of individual and general population whole body dose can be calculated from the PMS data, based on estimates of distance and population density near SLAC. Table 4 tabulates these results, and the model is discussed in Appendix B.

From this population estimate and the measured radiation dose near the site boundary, we can estimate both the average individual

dose and the population dose from SLAC operations. From 1974 data,

we arrive at 2040 persons who are included in the pool exposed to

1 mrem or more for the calendar year 1977. The man-rem dose was 3.9,

and the average individual dose was 1.9 mrem, or less than a 2%

increase over the average dose from natural background radiation.

Monitoring for Airborne Radioactivity

During 1977, 1.7 Ci of short-lived gaseous radioactivity was

released into the atmosphere from SLAC. Particulate samplers continue

to demonstrate that radioactivity in this form is not released from

SLAC. When corrected for dilution, this resulted in an average off-site

concentration of 2.5×10^{-11} Ci/m^3. This concentration is compared to

the CG for ^{41}Ar, which is 4×10^{-8} Ci/m^3, and is 0.06% of the

technical standard.

We emphasize that the model used to calculate off-site

concentration applies to the plume centerline, and is not corrected

for vertical and horizontal plume spread. Also, the model is not

corrected for wind direction or velocity. The estimate of off-site

concentration is, therefore, conservative and overestimates the actual

concentration at the site boundary by factors of 2-10.

Water Samples

Results of the environmental analysis for radioactivity found

in water appear in Table 3, and are reported in μCi/mℓ for gross beta

nonvolatile radioactivity. Analysis of tritium is performed on all

water samples, but is not reported because all environmental samples

analyzed exhibited a tritium content less than the minimum detectable

level of 3×10^{-6} μCi/mℓ.

Some of the peripheral wells (ground water) exhibit higher than normal concentrations of gross beta radioactivity and solids. This activity is due to naturally occurring radioactive potassium -40 and has been verified by flame photometry and gamma spectral analysis. The ^{40}K content in samples from W-1 are typically about $2 \times 10^{-7} \mu Ci/m\ell$, accounting for approximately 90% of the gross beta activity present. Only those samples whose concentrations are above $1 \times 10^{-7} \mu Ci/m\ell$ are corrected for ^{40}K. One well has exhibited some naturally occurring radioactive daughters of uranium.

NONRADIOACTIVE EFFLUENT

Waste Water

Waste water from SLAC is discharged in basically three directions:

(1) Sanitary sewer

(2) Storm sewer effluent is released to natural open ditches. The ditches conduct this water to San Francisquito Creek by surface flow or by seepage. Both these liquid effluents (1) and (2) eventually reach San Francisco Bay, about six miles to the east.

(3) About 40% of the water leaves the site as water vapor via the four cooling towers.

Typical amounts are 7.6×10^4 m^3/year to the sanitary sewer; 1.5×10^5 m^3/year to the creek; and 1.5×10^5 m^3/year evaporated. In addition, an average of 9×10^5 m^3/year fall as rain on the 472 acre site, also flowing to the creek. Thus, the SLAC effluent to the creek is diluted by an average factor of 6 by natural runoff.

Continuous sampling stations monitor the effluents of both the sanitary sewer and the storm sewer and have been in operation since the summer of 1971. The sanitary sewer sampling station is at Manhole No. 4, northeast of the Central Laboratory. All SLAC sanitary sewage flows through this point. A pump continuously samples the effluent at the rate of 5 mℓ per minute, which is stored in drums at that point. At the end of each quarter, the contents of the drums are mixed, and 1 liter samples are removed for chemical-radiological analysis.

It should be noted that the sampling rate is constant at all times regardless of the flow rate, and thus tends to give relatively greater weight to the effluent at lower flow rates when concentrations are likely to be higher. Therefore, this method of sampling is likely to lead to an overestimate rather than an underestimate of concentrations.

Sanitary Sewer Effluents

About 20% of SLAC's domestic water supply is released to the sanitary sewer; the remaining 80% leaves the site by evaporation or via San Francisquito Creek. The sanitary sewer outlet at the northeast corner of the site is connected to the Menlo Park Sanitary District. The releases are ordinary sanitary wastes, and the Menlo Park treatment plant discharges its treated wastes directly into San Francisco Bay.

The quantity of discharge for 1977 was $\sim 7.8 \times 10^4 \ m^3$. This amount is not unusual for a facility employing about 1000 people, and does not constitute a burden on the Menlo Park Sanitary District. The amount is rather constant the year around, and relatively insensitive to the accelerator operating cycle.

Storm Sewer Effluents

Water effluent discharged into the storm sewer is a combination of (1) cooling tower blowdown, (2) water runoff from SLAC landscaping irrigation, (3) rainwater runoff during the winter, and (4) miscellaneous uses, mainly once-through cooling of various small systems.

About a quarter of the 3.8×10^5 m^3 of water per year used at SLAC flows as waste water to the creek.

The largest potential source of chemical effluents discharged to San Francisquito Creek is cooling tower blowdown water, discharged to three natural open ditches. There are four towers of the induced-draft counterflow type. The primary system in all cases is a closed-loop, low-conductivity system. Tower 101 is located at the Central Utility Building, and provides cooling for laboratories and shops of the Campus Area. Tower 1201 is adjacent to the accelerator, and is meant to cool the injector, positron source, and the first accelerator mile. Tower 1202, also adjacent to the accelerator, cools the second mile, while Tower 1701, near the Beam Switchyard and Research Area, provides cooling for these areas.

The cooling tower water is chemically treated with silica and an organic algaecide compound. The blowdown water is basically source water whose solutes are concentrated by a factor of 4-6.

The cooling tower effluents are subject to control by the U.S. Environmental Protection Agency (EPA), under Discharge Permit No. CA0005541, issued September 30, 1973, and the State of California Regional Water Pollution Control Board (RWPCB). The EPA Discharge Permit states the maximum permissible amounts of concentrations of

total suspended solid, zinc (Zn), and net added phosphate (PO_4), as well as maximum temperature and permissible range of pH. It sets forth a monitoring schedule describing the types of sampling and minimum frequency of analysis. Each tower's effluent is analyzed separately, and quarterly reports are filed.

MONITORING RESULTS

Storm Sewer

Results of water monitoring efforts during 1977 appear in Table 5.

TABLE 2

ANNUAL PENETRATING RADIATION DOSE EQUIVALENT MEASURED NEAR SLAC's BOUNDARIES

1977

	Gamma			Neutron, mrem			%
	Background + Source	Background	Source	Background + Source	Background	Source[d]	Standard (b)
1	91	92	(a)	18.0	12.8	5.2	1.04
2	69	70	(a)	13.3	11.2	2.1	0.42
3	(c)	--	--	(c)	--	--	--
4	(c)	--	--	15.3	9.6	5.7	1.1
5	88	88	(a)	18.2	10.0	8.2	1.6
6	79	80	(a)	11.7	10.8	0.9	0.18
8	(c)	--	--	(c)	--	--	--

(a) Difference between background radiation and source contribution falls within normal variation of background values, and all are consistent with zero. Because these values are the difference between relative large numbers, each having fluctuations, negative numbers may occur.

(b) Standard from U.S. Energy Research and Development Administration Manual, Chapter 0524, 500 mrem per year.

(c) Station out of service.

(d) Owing to uncertainties in the neutron spectrum at the site boundary, and in the assignment of a quality factor, the neutron dose may be overestimated by a factor of 2 (see text).

TABLE 3

MEASURED RADIOACTIVITY CONTENT OF ENVIRONMENTAL SAMPLES
COLLECTED AT SLAC FROM JANUARY THROUGH DECEMBER 1977

Sample		Total Samples	Gross Beta Activity [a]			% Standard[1]
			Minimum	Maximum	Average	
Well Water,	μCi/ml	4	1.0×10^{-9}	1.5×10^{-7}	4.5×10^{-8}	(b)
Sanitary Sewer,	μCi/ml	2[c]	3.5×10^{-8}	3.7×10^{-8}	3.6×10^{-8}	(b)
Storm Sewer,	μCi/ml	1[c]			1.1×10^{-9}	(b)

(a) Includes ^{40}K and uranium.

(b) Does not apply: all samples consistent with background only.

(c) 1 aliquot sample from 2×10^{6} ml volume.

TABLE 4

SUMMARY OF RADIATION MEASUREMENTS
BY PATHWAY DURING CALENDAR 1977

Exposure Pathway	Maximum Annual Dose mrem	Percent of Standard	Man-Rem Estimate	
			SLAC	Background
Penetrating	8.2[a]	1.6	3.9	204
Water	(b)	(b)	(b)	50[d]
Airborne	<0.03	0.06	(c)	
Total	8.2	1.6	3.9	250

(a) Maximum measured value at PMS 5.

(b) Does not apply: all samples revealed background levels.

(c) Below significant levels.

(d) 25 mrem per person for internal does from natural radioactivity[5] or ~50 man-rem to population near SLAC.

TABLE 5

NONRADIOACTIVE EFFLUENT DISCHARGE MONITORING REPORT

(COOLING TOWER BLOW DOWN)

1977

Parameter	Sample Frequency	Range or MDL	Units	Stand	Cooling Tower 1201			Cooling Tower 1701			Cooling Tower 1202			Cooling Tower 101		
					Min(1)	Max	Aver	Min(1)	Max	Aver	Min(1)	Max	Aver	Min(I)	Max	Aver
FLOW	Continuous	Meter	Gal/d	NA	0	18500	1157	0	24710	649	0	18500	861	0	4840	25
pH	Continuous	0.1-14.0		6.0-9.0	7.8	8.6	8.22	7.3	8.9	8.11	7.8	8.5	8.26	7.3	8.9	8.21
TEMPERATURE	Daily	0.5	°F	85	52	76	63.5	58	76	64.4	55	77	64.3	56	73	64.5
ZINC	Monthly	0.001	mg/l	1.0	-	2.1	-	-	1.35	-	-	0.75	-	-	0.58	-
Phosphorous	Monthly	0.01	mg/l	0.1 Added	-	0.32	-	-	0.85	-		0.78	-	-	0.82	-
SETTLABLE SOLIDS	Monthly	>0.1	mg/l	0.1	-	2.5	-	-	0.3	-		0.1	-	-	0.1	-
Total Suspended Solids	Monthly	>0.1	mg/l	20	-	64	-	-	40	-		33	-	-	21.3	-

MDL is minimum detectable level of concentration analyses.

NA - not applicable

(1) Cooling Tower Blow Down is activated by a conductivity controller and this is intermittent.

APPENDIX A

Atmospheric Dispersion Model

In 1966, an independent evaluation of meterological regimes at
SLAC was performed.[4] From this study, an empirical mathematical model
was developed. The model that is used predicts the centerline
concentration very well, but overestimates the total dosage values.

$$\frac{X_P}{Q} = \frac{G}{u} \; \frac{X}{X_0}^{-1.75 + \frac{b(1-C)}{u}}$$

where X_P = centerline concentration (Ci/m^3)

Q = source strength (Ci/s)

$G = 8 \; m^{-2}$

u = mean wind speed (m/s)

X = distance from source (m)

$X_0 = 2 \; m$

C = fraction of sky covered by low clouds

$b = 0.5 \; m/s$ (day); $b = 1.2 \; m/s$ (night).

Figure 5 summarizes peak concentration per unit source strength as a
function of wind speed and atmospheric stability at a fixed distance
of 400 meters (roughly the distance from the source to SLAC's boundaries).
To characterize atmospheric stability, the degree of cloud cover is
indicated for day and nighttime regimes. This method is based upon
Pasquill's data for cloud expansion for various stability categories.

For a wind speed of 2 m/s atmospheric dilution factors—for
determining centerline concentrations—range between 2×10^{-5} and

1.5×10^{-3}. For purposes of estimateing radiation dose at the site boundary, neutral conditions are assumed, and a generally conservative dilution factor of 4.5×10^{-4} s/m^3 is used in calculation of average concentration at the site boundary. (See Figure 5, Curve 1.0.)

APPENDIX B

Model for Potential Dose Assessment

According to Chapter 0513 "Effluent for Environmental Monitoring and Reporting," an assessment of whole body man-rem dose to the general population within 50 miles (80 km) of SLAC is required. Our site boundary dose due to accelerator operation is detectable, and is <20 mrem per year from penetrating radiation. Integrating a population dose of small values out to 50 miles becomes an exercise in numerical analysis that results in questionable dose estimates. This is true because of the questionable assumptions that must be made to explain the behavior of neutrons at large distances from the source. We have, therefore, modified the distance term to include individual annual doses down to 1 mrem, which corresponds to a distance of <1 mile from a central point representative of the source of neutrons. The 1 mrem value is approximately 1% of the total nautral background dose, and any further extrapolation is unjustified because the difference in population dose from natural background and SLAC operations cannot be reasonably determined.

There are three major pathways leading to human exposure: (1) airborne, (2) food chain, and (3) direct exposure to penetrating radiation. Of the three major pathways listed above, only direct exposure to penetrating radiation is of any measurable significance from SLAC operations. The source of this exposure is from neutrons resulting from the absorption of high energy electrons and photons in the experimental area creating energetic particles, some of which escape from the heavily shielded areas.

In order to make an accurate and realistic assessment of

radiation exposure to the public at low doses, it is necessary that

exposure from the natural radiation environment be known. This is true

because the instruments used for this purpose respond to natural

radiation sources as well as man-made sources, and the portion due to

natural radiation must be subtracted from the raw results. The

population exposure assessments appearing in this document are, in all

cases, overstatements of the true impact. Hence, the resulting values

are representative of an upper limit of the possible range.

While the annual neutron dose from accelerator operations at the

site boundary is measurable, it amounts to <25% of the total annual

individual dose from natural background radiation. According to an

EPA report, the average dose from cosmic, terrestial, and internal

radiation in California is 125 mrem.[5] For purposes of comparison, we

have rounded this number off to 100 mrem.

Another quantity of interest is the population dose in units of

man-rem. This is simply the product of the average individual dose

and the total population exposed. For example, if there are 2000

people exposed to an average annual background does of 0.1 rem (100 mrem),

then the population dose is 0.1 x 2000, or 200 man-rem from natural

background radiation. The annual variation of exposure to natural back-

ground radiation may vary by $\pm 20\%$, largely caused by the difference of

naturally occurring uranium, thorium, and potassium present in the ground

and in building materials where people live and work. This value is

also affected by weather conditions which may increase or decrease the

amount of randon/thoron present in the atmosphere at any given time.

There are two major problems associated with this determination that affect overall accuracy of the measurement. First, the conversion of neutron flux to dose requires that the spectrum of neutrons at the measurement point be known. In order to characterize the SLAC neutron spectrum, a much larger source term is required. Since the quality factor (QF) is a function of energy, we have selected a QF of 10 as best characterizing the energy spectrum produced by SLAC. We feel that this is an overestimate by a factor of approximately 2. This degree of conservation overestimates the actual dose, also by a factor of approximately 2. Until a useful experiment can be performed, with neutron yields of sufficient intensity, the QF cannot be determined with any better precision.

A second problem with this sort of extrapolation is the behavior of neutrons at large distances. Most of the high energy accelerator laboratories have made measurements and have derived formulas for predicting this behavior.[8] Unfortunately, all such measurements are unique to each facility because of deisgn differences, type of machine, and surrounding topography. Here, again, we have chosen a conservative formula for calculating the dose at distances other than the point of measurement. Lindebaum[6] gave a method for evaluating skyshine neutrons which was later verified by Ladu et al.[7] using Monte Carlo techniques. Lindenbaum approximated the falloff by $e^{-R/\lambda}/R$, where R is in feet, and λ = 830 feet. This equation fits the SLAC data fairly well, and is the one used to predict doses beyond our measuring station (Figure 6). In order to derive a correction for large distances unique for SLAC, we will

need a much larger intensity to determine a more precise correction for distance. We feel that the methods used and reported in this document may overestimate the true population dose by at least a factor of 2 or greater.

The population activity close to SLAC, i.e., within 1 mile, is a mixture of commerce and residential dwellings. The occupancy factor—the proportion of time throughout the year that these structures are occupied—is assumed to be 1/4 for business activities, and 1.0 for private dwellings. The number of people is estimated for each type of structure, multiplied by the occupancy factor, and summed to estimate the total population that might be continuously present. (See Figure 4.)

REFERENCES

1. US. Energy Research and Development Administration Manual, Chapter 0524, Standards for Radiation Protection.

2. W. R. Nelson, "Radioactive Ground Water Produced in the Vicinity of Beam Dumps," Stanford Linear Accelerator Center Report No. SLAC-TN-65-16 (July 1965).

3. Recommendations of the International Commission on Radiological Protection, Publication 2 (Pergamon Press, London, 1959).

4. J. A. Murray, L. M. Vaughn, anf R. W. McMullen, Atmospheric Transport and Diffusion Characteristics for Selected Daytime Meterological Regimes at SLAC, Memorandum Report No. 326-1, Metronics Aerosol Laboratory, 21 December 1967.

5. "Estimates of Ionizing Radiation Doses in the United States, 1960-2000," U.S. Environmental Protection Agency Report No. ORP/CSD 72-1 (1972).

6. S. J. Lindenbaum, "Shielding of High Energy Accelerators," Int. Conf. on High Energy Accelerators, Brookhaven, 1961.

7. M. Ladu et al., "A Contribution to the Skyshine Study," Nucl. Instr. and Meth. 62, 51 (1968).

8. T. M. Jenkins, "Accelerator Boundary Doses and Skyshine," Health Physics 27, 251 (1974).

External Distribution:

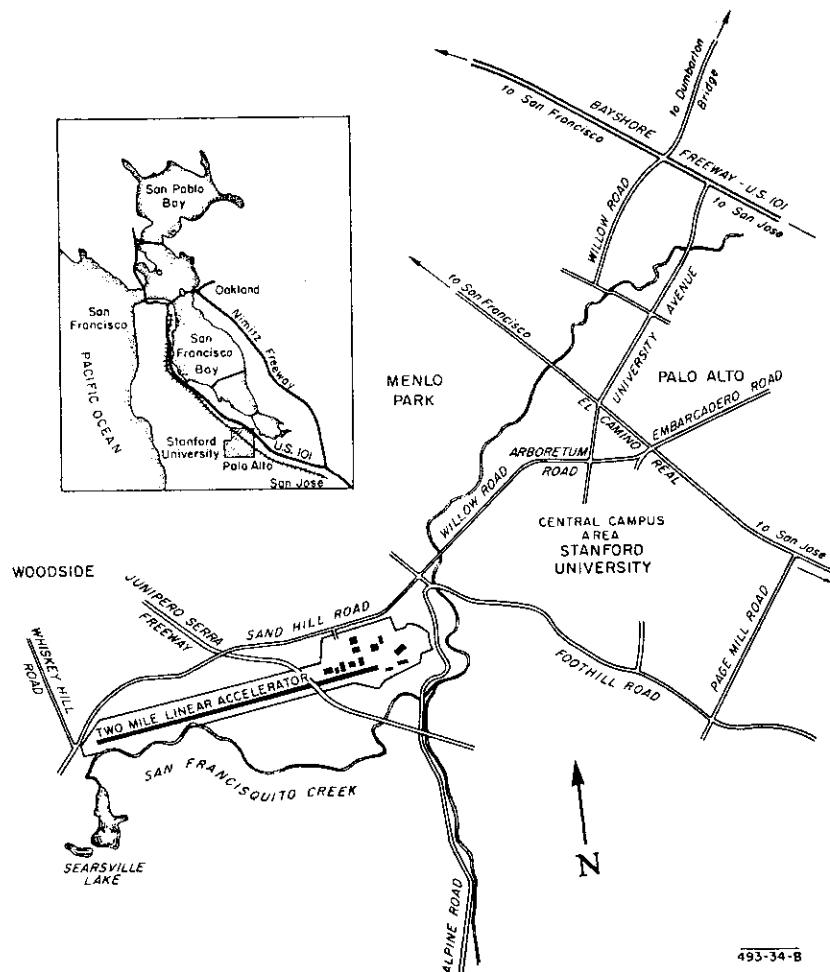|	
Gordon Facer, MA	DOE-HQ
Andrew J. Pressesky, RDT	DOE-HQ, Asst. Dir. for Nuclear Safety
James Miller, BER	DOE-HQ
DOE-HQ Library Attn: Charles Sherman	
Jack L. Brown	BRH-CALIF
Regional Administrator	EPA Region IX

FIG. 1--Site location relative to Stanford University
and surrounding communities.

FIG. 2--Air view of SLAC site showing the two-mile
accelerator, the research facility, and the
principal laboratories and shops.
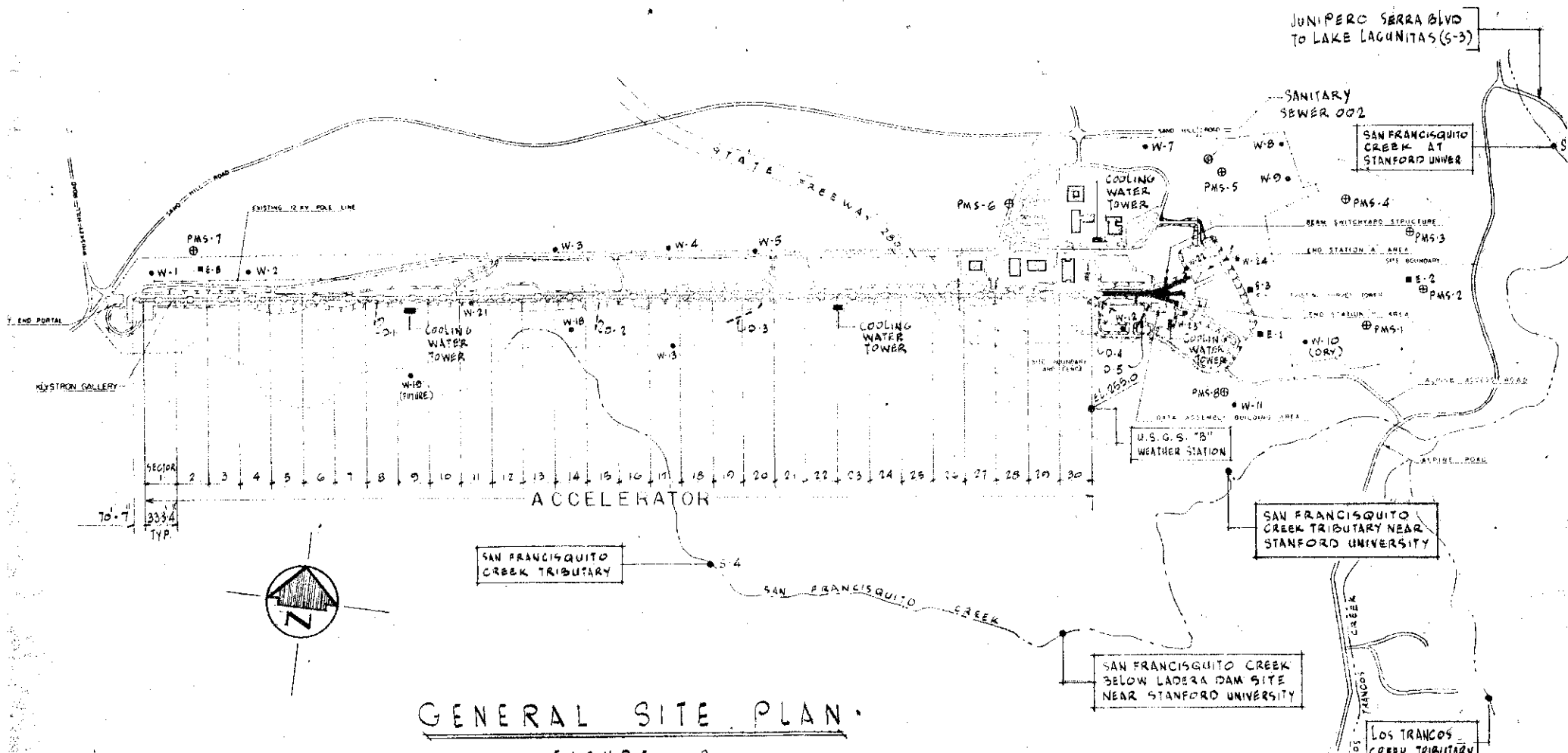
GENERAL SITE PLAN

FIGURE 3

SCALE: 1" = 1000'

SAMPLE STATION LEGEND

● W PERIPHERAL WELLS
● S SURFACE WATER
■ E ENVIRONMENTAL PLOTS
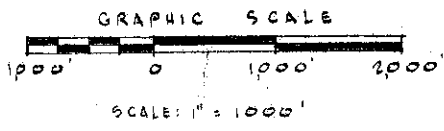⊕ PMS PERIPHERAL MONITORING STATIONS
---D SUBDRAINS
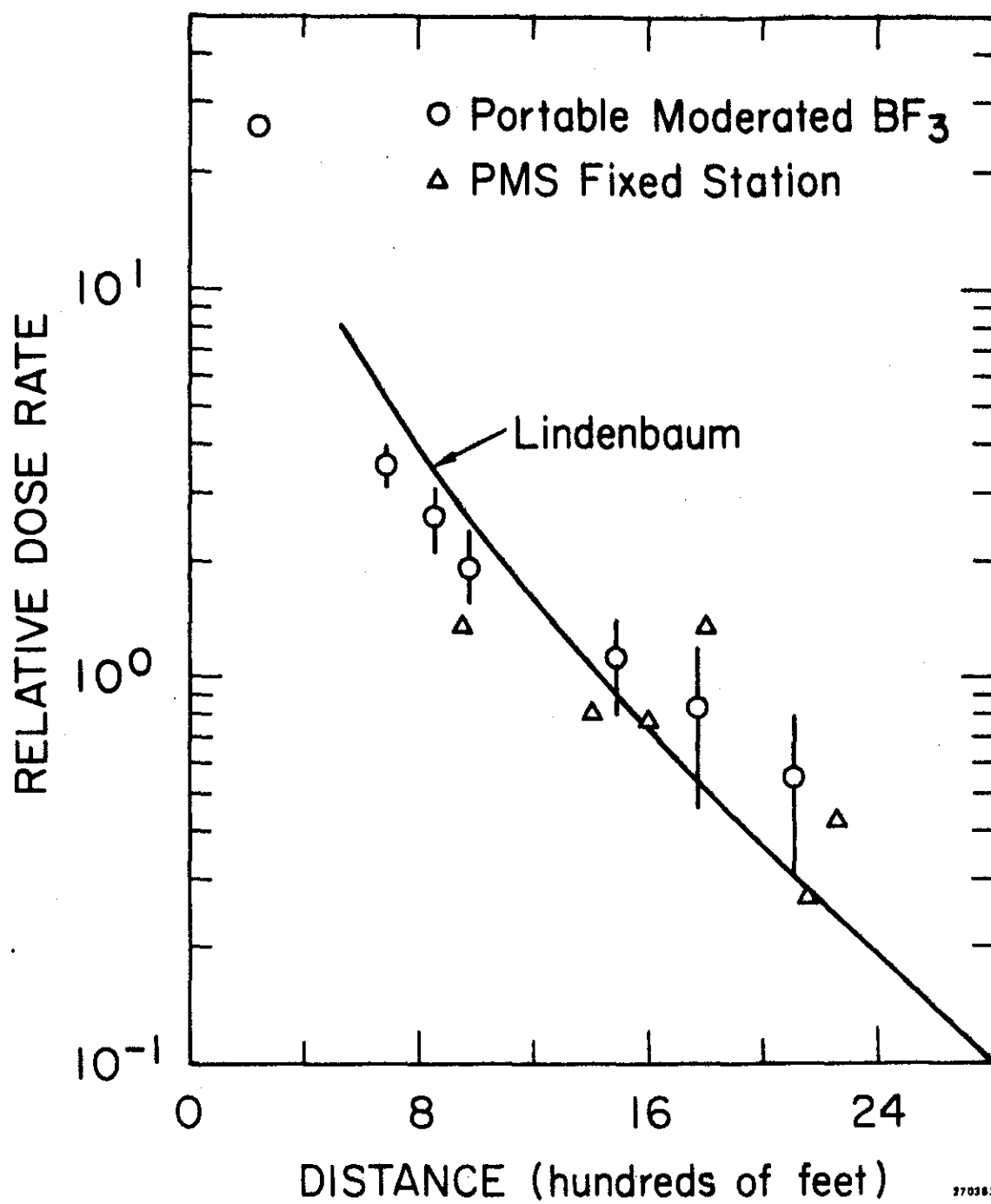
FIG. 4--SLAC research yard and surrounding community.

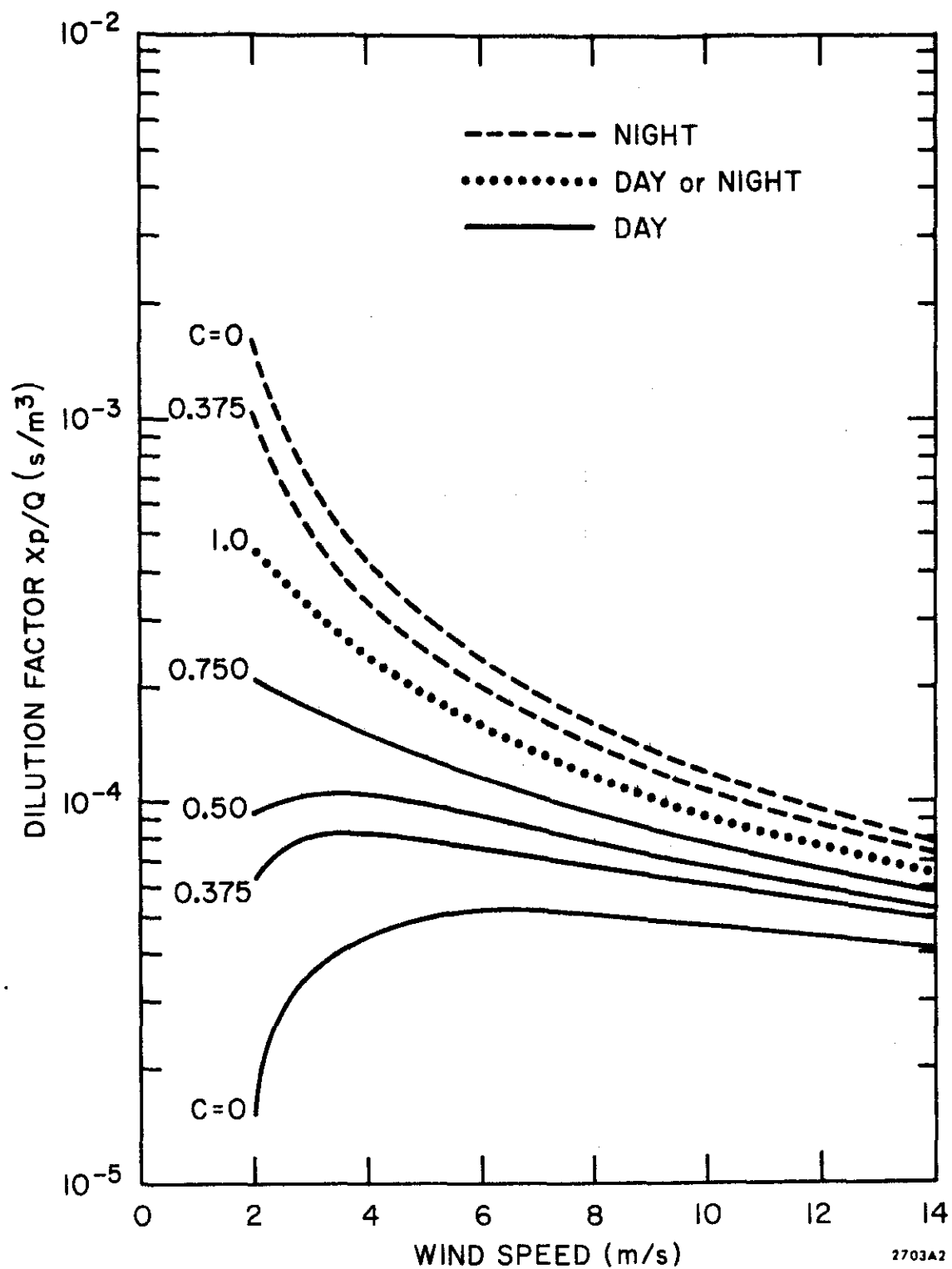FIG. 6--Measurements made along a line between ESA and site boundary.

FIG. 5—Centerline dilution factor for various atmospheric conditions as a function of wind speed.